

An Exploration on Greenhouse Gas and Ammonia Production by Insect Species Suitable for Animal or Human Consumption

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Abstract

Background: Greenhouse gas (GHG) production, as a cause of climate change, is considered as one of the biggest problems society is currently facing. The livestock sector is one of the large contributors of anthropogenic GHG emissions. Also, large amounts of ammonia (NH₃), leading to soil nitrification and acidification, are produced by livestock. Therefore other sources of animal protein, like edible insects, are currently being considered.

Methodology/Principal Findings: An experiment was conducted to quantify production of carbon dioxide (CO₂) and average daily gain (ADG) as a measure of feed conversion efficiency, and to quantify the production of the greenhouse gases methane (CH₄) and nitrous oxide (N₂O) as well as NH₃ by five insect species of which the first three are considered edible: *Tenebrio molitor*, *Acheta domesticus*, *Locusta migratoria*, *Pachnoda marginata*, and *Blaptica dubia*. Large differences were found among the species regarding their production of CO₂ and GHGs. The insects in this study had a higher relative growth rate and emitted comparable or lower amounts of GHG than described in literature for pigs and much lower amounts of GHG than cattle. The same was true for CO₂ production per kg of metabolic weight and per kg of mass gain. Furthermore, also the production of NH₃ by insects was lower than for conventional livestock.

Conclusions/Significance: This study therefore indicates that insects could serve as a more environmentally friendly alternative for the production of animal protein with respect to GHG and NH₃ emissions. The results of this study can be used as basic information to compare the production of insects with conventional livestock by means of a life cycle analysis.

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Introduction

Production of greenhouse gasses (GHG) is considered as an important cause of climate change [1,2]. The most important GHGs are carbon dioxide (CO₂), methane (CH₄) and nitrous oxide (N₂O). Since the end of the 18th century the atmospheric carbon-dioxide concentration has increased by 30% and CH₄ concentrations by 50% [3]. CH₄ and N₂O have considerably greater global warming potentials (GWPs) than CO₂. By assigning CO₂ a value of 1 GWP, the warming potentials of these other gases can be expressed on a CO₂-equivalent basis: CH₄ has a GWP of 25, and N₂O has a GWP of 298 [1]. The relative contribution of CO₂ equivalents (CO₂ eq.) of the livestock sector is large, amounting up to 18% of total anthropogenic GHG emissions [2]. Based on a Life Cycle Analysis (LCA) that takes the entire production process of animal products into account, the global contribution to GHG emissions by the animal sector are: 9% for CO₂ (fertilizer production for feed crops, on-farm energy expenditures, feed transport, animal product processing, animal

transport, and land use changes), 35–40% for CH₄ (enteric fermentation in ruminants and from farm animal manure) and 65% for N₂O (farm manure and urine) [2]. Direct CO₂ production through respiration is not relevant when determining the impact of GHGs as respiration by livestock is not considered a net source of CO₂ [2]. The respired carbon, which comes from the feed, was first taken up from CO₂ in the air and stored in an organic compound during the production of the feed. However, the ratio between body growth realised and CO₂ production is an indicator of feed conversion efficiency and thereby a relevant indicator for the environmental impact [4].

Livestock is also associated with environmental pollution due to ammonia (NH₃) emissions from manure and urine, leading to nitrification and acidification of soil [5]. Although not considered a GHG, NH₃ can indirectly contribute to N₂O emission [2], as conversion takes place by specialized soil bacteria [6]. Livestock is estimated to be responsible for 64% of all anthropogenic NH₃ emissions [2]. The main source of gaseous NH₃ is bacterial fermentation of uric acid in poultry manure [7,8] and bacterial

fermentation of urea in mammals [9]. Besides these environmental problems the livestock sector faces challenges regarding resistance to antibiotics, zoonosis and animal welfare [10].

All these problems together illustrate the need to find alternatives for conventional sources of animal protein. Mini-livestock, for instance edible insects, have been suggested as an alternative source of animal protein [11]. Production of animal protein in the form of edible insects supposedly has a lower environmental impact than conventional livestock [12,13,14]. When evaluating the total environmental impact of animal protein production, a LCA, in which all production factors are taken into account, is needed. Differences in environmental impact in a LCA can be explained mainly by three factors: enteric CH₄ emissions, feed conversion efficiencies and reproduction rates [4].

Before performing a LCA, it is necessary to know the GHG production by edible insects. This information is lacking in literature. Therefore, in this study we experimentally quantified the direct production of the GHGs CH₄ and N₂O for five insect species. CO₂ production and average daily gain (ADG) were quantified to provide an estimation of feed conversion efficiency. Additionally, NH₃ emissions were quantified. The results of this study represent a quantification of the insect physiological contribution to GHG production by insects and can in turn be used to create a LCA for insect-derived products.

Materials and Methods

2.1 Animals and housing

Five insect species were studied: fifth larval stage mealworms *Tenebrio molitor* L. (Coleoptera: Tenebrionidae), fifth and sixth nymphal stage house crickets *Acheta domesticus* (L.) (Orthoptera: Gryllidae), third and fourth stage nymphs of migratory locusts *Locusta migratoria* (L.) (Orthoptera: Acrididae), third larval stage sun beetles *Pachnoda marginata* Drury (Coleoptera: Scarabaeidae) and a mix of all stages of the Argentinean cockroach *Blaptica dubia* (Serville) (Dictyoptera: Blaberidae). Currently, *T. molitor*, *A. domesticus* and *L. migratoria* are considered edible, while *P. marginata* and *B. dubia* are not. The latter two species were included since they are a potential source of animal protein, for instance by means of protein extraction. These two species can be bred in large numbers with little time investment and are able to utilise a wide range of substrates as feed [15,16].

Per species three to six repetitions were conducted each for a period of three days. Animals were housed per species in two cages or containers per respiration chamber. These containers were placed in one of two, identical, open circuit climate respiration chambers measuring 80*50*45 cm, with a total volume of 265 L [17]. Within these climate respiration chambers, *T. molitor* and *P. marginata* were housed in two stacked plastic containers

(50*30*8.7 cm). The three other species were housed in metal wire cages (45*37.5*41 cm; mesh width 1 mm) with a glass cover plate. To increase surface area for *A. domesticus* and *B. dubia*, hollow plastic tubes (20 cm long and 3 cm in diameter), were stacked to a height of 30 cm in the wired cages, while for *L. migratoria*, two V-shaped-folded metal screens (70*15 cm) were entered per cage. Humidity, temperature, and day length were based on rearing conditions used by commercial insect rearing companies (Table 1). All animal masses reported are averages of fresh mass per cage. The starting and final animal mass per cage are provided in Table 1.

2.2 Diet

Food was provided for each species at the beginning of each repetition, except when mentioned otherwise.

Tenebrio molitor larvae were reared in 300 g mixed grain substrate (wheat, wheat bran, oats, soy, rye and corn, supplemented with beer yeast) with on top pieces of carrot ($\pm 15*2$ cm) weighing a total average of 637 g per repetition.

Acheta domesticus was provided with chicken mash (501 g) with carrot pieces (784 g) on top for each repetition.

Locusta migratoria was provided with wheat bran (70 g; Arie Blok Animal Nutrition, Woerden, The Netherlands) in a metal bowl at the beginning of each repetition. Fresh Perennial ryegrass (*Lolium perenne*) was provided daily (463 g in three days). The grass was grown by Unifarm, Wageningen University and Research centre, Wageningen, The Netherlands.

P. marginata larvae were kept in a peat moss substrate (2.0 kg per respiration chamber) in which chicken mash (285 g) was mixed at the beginning of each three-day repetition. Pieces of carrot ($\pm 15*2$ cm) with an average total mass of 161 g per repetition were put on top of the substrate.

B. dubia was provided with a chicken mash diet (199 g) and carrots (559 g), fresh carrot being added during the repetitions.

Peat moss, chicken mash, and carrots, offered to *A. domesticus*, *P. marginata* and *B. dubia* were provided by Kreca V.O.F, Ermelo, The Netherlands. The carrots and mixed grains substrate offered to *T. molitor* were provided by Insectra, Deurne, The Netherlands.

2.3 Gas measurements

During the experiment concentrations of CO₂ and CH₄ were measured every 9 min in the ingoing and outgoing air stream of the respiration chambers. The difference in CO₂ and CH₄ concentrations between ingoing and outgoing air thus represents the total production of CO₂ and CH₄ of insects, feed, and substrate. The exact air volumes were measured with a calibrated Schlumberger G1.6 dry gas meter and corrected for measured air temperature and pressure. CO₂ and CH₄ concentrations were

Table 1. Mean values and standard deviations of temperature, humidity, ventilation, hours of light per day and average start and final weight for five insect species.

	<i>Pachnoda. marginata</i>	<i>Tenebrio molitor</i>	<i>Blaptica dubia</i>	<i>Acheta domesticus</i>	<i>Locusta migratoria</i>
Temperature (°C)	28.0±0	25.0±0	28.0±0	28.0±0	32.0±0
Humidity (%)	84.3±3.3	79.8±0.2	70.0±0.0	69.9±0.1	69.7±0.2
Ventilation (L/min)	6.46±2.06	6.82±1.31	5.16±0.05	11.18±1.80	4.98±0.39
Hours of light per day	0	0	12	12	12
Start weight (kg)	0.99	0.91	1.10	0.96	0.08
Final weight (kg)	1.10	1.10	1.28	1.17	0.13

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measured in dried gas. Gas was dried in a +2°C dew-point cooler. Nondispersive infrared analyzers were used to measure CO₂ (type Uras 3G, Hartmann and Braun, Frankfurt, Germany) and CH₄ (type Uras 10E, Hartmann and Braun, Frankfurt, Germany). The refreshed air volume was set so that CO₂ levels did not exceed 1%. From each climate respiration chamber, as well as from the incoming air, an air sample was taken for N₂O analysis after 24, 48, and 72 h with a 60 ml syringe. The syringes were sealed by a shutoff valve and stored at 20°C until analysis (within 48 h). The N₂O concentration was analysed by a gas chromatograph (CE instruments GC8000 Top, Interscience, Breda, The Netherlands) using a Haysep Q 80–100 mesh 2 m×1/8" SS column, at a constant temperature of 60°C. N₂O was detected with an electron capture detector (ECD). Injection volume was 5.0 ml in a fixed loop.

NH₃ concentrations in the climate respiration chambers were determined twice daily (at 12.00 and 24.00 h) by means of a gas detection tube system (Kitagawa, type AP-20; Komyo rikagaku kogyo, Tokyo, Japan; type 105 NH₃ gas detector tubes with a range of 1–20 ppm).

2.4 Calculations

Production of N₂O was calculated by subtracting the N₂O concentration from the incoming air from that in the outgoing air. These differences were then used in a formula adapted from Wheeler et al (2003) [18]:

ER = Emission rate of N₂O = $[N_2O]$ change (ppm×10⁻⁶)×VV (m³/day)×44 (g/mol)/0.0224 (m³/mol), where VV = ventilation volume of air in a specified time period. The average concentration difference of the three samples taken during the three-day period was used to determine the average N₂O production in a repetition.

The formula used by Wheeler (2003) was also used for the calculation of NH₃ production. A molecular mass of 17 was used and instead of a difference in concentration, the measured concentration was used, leading to a slight overestimation of the actual NH₃ production (between 0 and 0.1 mg/kg BM/day).

CO₂ equivalents were calculated by adding the multiplications of the produced amounts of CH₄ and N₂O with their global warming potential; 25 for CH₄, and 298 for N₂O [1].

Mean body mass was calculated by averaging the body mass at the start of the experiment and the body mass at the end of the experiment. Average daily gain (ADG) was calculated as follows:

$((\text{End mass} - \text{Start mass})/\text{Start mass})/3 \times 100\%$, in which 3 is the number of days the experiment was running.

The ratio between CO₂ production per unit biomass per day and ADG gives an indication of the feed conversion efficiency, in which higher values indicate lower efficiencies.

To determine CO₂ production from feed and substrate, all feeds were independently tested in the same respiration chambers, without the animals. A linear time course of consumption was assumed and CO₂ production was recalculated to kg of live insect.

2.5 Statistics

The N₂O and NH₃ assay data were subjected to a two-way analysis of variance (ANOVA) with species and time of sampling (24, 48, or 72 h) as fixed factors to determine whether the time of sampling had an effect. No significant effect of the time of sampling was found for N₂O (Pillai's trace: F = 1.467, P = 0.199). Therefore, the average of the three samples taken during the 3-day trial period was used to determine the change per repetition and to calculate total production. However, NH₃ production was significantly affected by the time of sampling (day or night; Pillai's trace: F = 4.065, P = 0.019) and the day of the repetition (first, second or third; Pillai's trace: F = 17.170, P < 0.001). CO₂ and CH₄ production for all five species were analyzed by means of a one way analysis of variance (ANOVA) followed by a Tukey post hoc test. Statistical analysis of all data was done by means of SPSS 15.0.

Results

Production of CO₂ is expressed per kilogram of mean live body mass (BM) per day (24 hours) and per kilogram of mass gain (Table 2) and the average daily gain (ADG) is reported (Table 2). Production of CH₄, N₂O, CO₂ equivalents, and NH₃, are expressed per kilogram of mean live body mass (BM) per day (Table 3) and per kilogram of mass gain (Table 4).

3.1 ADG and CO₂ production

ADG varied between 4.0% (*P. marginata*) and 19.6% (*L. migratoria*) with the three other species having an ADG of 6–7%. CO₂ production among the five insect species differed significantly and ranged from 19 (*B. dubia*) to 110 (*L. migratoria*) g per kg BM/day. Also, the CO₂ production per kg of metabolic weight (i.e. the weight of metabolically active body tissue) differed greatly between

Table 2. CO₂ production (average ± standard deviation) per kilogram of bodymass per day, per kg of mass gain and average daily gain for five insect species, pigs and beef cattle.

Species	CO ₂ (g/kg BM/day)	CO ₂ (g/kg mass gain)	ADG (%)
<i>Pachnoda marginata</i> (n = 4)	50 ± 22 ^a	1,539 ± 518 ^a	4.0 ± 2.1% ^a
<i>Tenebrio molitor</i> (n = 4)	61 ± 9 ^b	1,031 ± 349 ^b	7.3 ± 2.5% ^b
<i>Blaptica dubia</i> (n = 3)	19 ± 3 ^c	337 ± 51 ^c	6.1 ± 0.7% ^c
<i>Acheta domesticus</i> (n = 4)	68 ± 10 ^d	1,468 ± 971 ^a	7.2 ± 3.4% ^b
<i>Locusta migratoria</i> (n = 6)	110 ± 21 ^e	734 ± 119 ^d	19.6 ± 2.1% ^d
Pigs	21.6–29.6	865–1,194	3.2 ± 0.53%
Beef cattle	5.3–7.0	2,835	0.3 ± 0.07%

BM = Body Mass;

ADG = Average daily gain;

Reported values for pigs and beef cattle were obtained from: [5] Aarnink et al., 1995; [49] Groot Koerkamp et al., 1998; [52] Demmers et al., 2001; [50] Nicks et al., 2003; [59] Beauchemin & McGinn, 2005; [48] Cabaraux et al., 2009 and [53] Harper et al., 2009. Mean values bearing different superscripts in a column differ significantly (P < 0.05).

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Table 3. CH₄, N₂O, CO₂ eq. and NH₃ production (average ± standard deviation) per kilogram of bodymass per day for five insect species, pigs and beef cattle.

Species	CH ₄ (g/kg BM/day)	N ₂ O (mg/kg BM/day)	CO ₂ eq. (g/kg BM/day)	NH ₃ (mg/kg BM/day)
<i>Pachnoda marginata</i> (n = 4)	0.16±0.085 ^a	0.0±0.03 ^a	4.00±2.13 ^a	0.1±0.16 ^a
<i>Tenebrio molitor</i> (n = 4)	0.00±0.002 ^b	1.5±0.13 ^b	0.45±0.04 ^b	0.0±0.09 ^a
<i>Blaptica dubia</i> (n = 3)	0.08±0.021 ^c	0.3±0.24 ^a	2.12±0.57 ^c	3.0±1.63 ^b
<i>Acheta domesticus</i> (n = 4)	0.00±0.002 ^c	0.1±0.13 ^a	0.05±0.04 ^b	5.4±3.40 ^c
<i>Locusta migratoria</i> (n = 6)	0.00±0.017 ^c	8.0±13.50 ^b	2.37±4.02 ^c	5.4±1.65 ^c
Pigs	0.049–0.098	2.7–85.6	2.03–27.96	4.8–75
Beef cattle	0.239–0.283	N/A	5.98–7.08	14–170

BM = Body Mass;

N/A = Not Available;

Reported values for pigs and beef cattle were obtained from: [5] Aarnink et al., 1995; [49] Groot Koerkamp et al., 1998; [52] Demmers et al., 2001; [50] Nicks et al., 2003; [59] Beauchemin & McGinn, 2005; [48] Cabaraux et al., 2009 and [53] Harper et al., 2009. Mean values bearing different superscripts in a column differ significantly (P<0.05).

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species (Table 5). CO₂ production expressed per kg of mass gain was intermediary for *L. migratoria* due to the high ADG. Still, the CO₂ production of *L. migratoria* per kg of mass gain was more than double the production of CO₂ by *B. dubia*. *Pachnoda marginata* had the highest production of CO₂ per kg of mass gain (1,539 g/kg), which was more than double the amount of *L. migratoria*.

3.2 CH₄

Production of methane was detected for *P. marginata* and *B. dubia*, but not for the three other species. *Pachnoda marginata* produced more than three times as much CH₄ per kg of mass gain than *B. dubia* (4.9 vs 1.4 g). This difference was caused by a higher production of CH₄ per kg BM (0.16 g vs 0.08 g) and a lower ADG (4.0% vs 6.1%).

3.3 N₂O

N₂O was produced only in significant amounts by *T. molitor* and *L. migratoria* (1.5 and 8.0 mg/kg BM/day, respectively). Production of N₂O by *L. migratoria* per kg BM was more than 5-fold the production by *T. molitor*, this difference decreased to almost 2.5-fold when expressed per kg of mass gain, due to a much higher ADG of *L. migratoria*.

3.4 NH₃

NH₃ was produced by *A. domesticus*, *L. migratoria*, and *B. dubia* (3.0–5.4 mg/kg BM/day), and ranged from 36–142 mg/kg of mass gain (Table 3 and 4). Significant differences (Pillai's trace: F = 4.065, P = 0.019) between daytime (12.00) and night-time (24.00) NH₃ emission levels were found for *A. domesticus* (6.4 and 4.4 mg/kg BM/day), *L. migratoria* (5.6 and 3.9 mg/kg BM/day), and *B. dubia* (3.4 and 2.6 mg/kg BM/day).

Discussion

Insects, being poikilotherms, do not use their metabolism to maintain a body temperature within narrow ranges, contrary to homeothermic animals. This is expected to result in higher feed conversion efficiencies. CO₂ production related to growth, has an inverse relationship with feed conversion efficiency in a given situation. CO₂ production by insects depends on the species, stage of development [19,20], temperature [21], feeding status [22], and on activity level [23,24]. A production of 37 g CO₂/kg BM/day was reported for *Anabrus simplex* (Orthoptera, Tettigoniidae), 40 g CO₂/kg BM/day for the locust *Schistocerca americana* (Orthoptera; Acrididae) [25] and 94 g/kg BM/day for adult *Tribolium castaneum* (Coleoptera; Tenebrionidae) [26]. All five species in the current

Table 4. CH₄, N₂O, CO₂ eq. and NH₃ production (average ± standard deviation) per kilogram of mass gain for five insect species, pigs and beef cattle.

Species	CH ₄ (g/kg mass gain)	N ₂ O(mg/kg mass gain)	CO ₂ eq. (g/kg mass gain)	NH ₃ (mg/day/kg mass gain)
<i>Pachnoda marginata</i> (n = 4)	4.9±1.96 ^a	1.03±1.06 ^a	121.86±49.09 ^a	3±4.8 ^a
<i>Tenebrio molitor</i> (n = 4)	0.1±0.03 ^b	25.5±7.70 ^b	7.58±2.29 ^b	1±2.0 ^a
<i>Blaptica dubia</i> (n = 3)	1.4±0.30 ^c	5.7±4.05 ^a	37.54±8.01 ^c	54±31.1 ^a
<i>Acheta domesticus</i> (n = 4)	0.0±0.09 ^b	5.3±6.05 ^a	1.57±1.80 ^d	142±184.5 ^b
<i>Locusta migratoria</i> (n = 6)	0.0±0.11 ^b	59.5±104.8 ^c	17.72±31.22 ^e	36±10.8 ^a
Pigs	1.92–3.98	106–3457	79.59–1,130	1140–1920
Beef cattle	114	N/A	2,850	N/A

BM = Body Mass;

N/A = Not Available;

Reported values for pigs and beef cattle were obtained from: [5] Aarnink et al., 1995; [49] Groot Koerkamp et al., 1998; [52] Demmers et al., 2001; [50] Nicks et al., 2003; [59] Beauchemin & McGinn, 2005; [48] Cabaraux et al., 2009 and [53] Harper et al., 2009. Mean values bearing different superscripts in a column differ significantly (P<0.05).

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Table 5. CO₂ production (g) per kilogram of metabolic weight per day for five insect species, pigs and beef cattle based on Kleiber's law ($B = aM^b$).

Species	b = 0.67	b = 0.75	b = 0.82
<i>Pachnoda marginata</i> (n = 4)	7	11	17
<i>Tenebrio molitor</i> (n = 4)	3	7	12
<i>Blaptica dubia</i> (n = 3)	2	4	6
<i>Acheta domesticus</i> (n = 4)	4	8	14
<i>Locusta migratoria</i> (n = 6)	9	17	29
Pigs	63	50	41
Beef cattle	50	31	21

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study had a fairly high production of CO₂. This might to a large extent be explained by *ad libitum* feeding during the experiment that has been reported to increase oxygen consumption fivefold [22]. Reported CO₂ production for inactive, unfed, Tenebrionid adults ranged between 5.4–13.3 g/kg BM/day [27], which is 5–10 times lower than observed for *T. molitor* in this experiment. This can partially be explained by the locomotory activities of *T. molitor* larvae in this experiment [37]. Furthermore, growing larvae are expected to have a higher CO₂ production than adults. The range of CO₂ production for *T. molitor* is comparable to the factorial metabolic scope reported for tiger beetles (*Cicindela* spp: Coleoptera; Cicindelidae) of 6.1–16.5 [28].

Size differences in animals account for a difference in metabolic rate, and thereby CO₂ production. The relation between metabolic rate (B) and body mass (M) was described by Kleiber [29] as $B = aM^b$, in which a is a constant and b = 0.75. The value of b has been much debated since [30,31,32]. For poikilotherms values between 0.67 and 1.0 have been reported and a comparison of several arthropod species suggested b approximates 0.82 [33,34]. The value chosen for b has a large impact on the metabolic weight and thereby the calculated CO₂ production (Table 5). Applying b = 0.75 for pigs and beef cattle and b = 0.82 for insects, resulted in a lower CO₂ production based on metabolic weight for the studied insect species (Table 5). For *L. migratoria* CO₂ production was only slightly lower than for beef cattle, however, for the other four species production was between 18% and 54% of that for beef cattle and between 11% and 34% of the CO₂ production of pigs.

The CO₂ production per kg BM of insect species investigated in this study was higher than for pigs or cattle (Table 3). This concurs with Prothero *et al.* (1979) [35], who reported a higher oxygen consumption per kg of BM for insects than for mammals, assuming the respiratory quotient (CO₂ production/O₂ consumption) has similar values (0.7–1.0) for both animal groups. However, the CO₂ production per kg of mass gain for the five insect species in the current study (337–1,539 g/kg) was either 39% (minimum values) or 129% (maximum values) when compared with pigs (865–1,194 g/kg) and much lower (12%–54% respectively) than cattle (2,835 g/kg). Therefore, CO₂ production per kg of mass gain suggests higher feed conversion efficiencies for insects than for mammalian livestock. These results concur with those of other authors [13,14,36,37].

A similar trend was visible for ADG; the ADG for the five insect species studied was 4.0–19.6%, the minimum value of this range being close to the 3.2% reported for pigs, whereas the maximum value was 6 times higher. Compared to cattle (0.3%), insect ADG values were much higher. In general, the rate of ADG depends,

amongst others, on life phase. Therefore, where available, literature data on growing animals were used. The fundamental biological differences in growth and development processes between pigs and cattle and the studied insects impeded further synchronization.

CH₄ production for the species studied was in agreement with Hackstein and Stumm (1994) [38]; for insects, only representatives of cockroaches, termites, and scarab beetles produce CH₄. This originates from bacterial fermentation by methanobacteriaceae in the hindgut [39].

We found large variability for the N₂O emission rates. Earlier studies in laying hens using a similar method for determining N₂O production, concluded that production was either negligible or undetectable [7,40]. However, other authors [41,42] determined a production of 28 mg N₂O/kg BM/day and 52 mg N₂O/kg BM/day, respectively, indicating the difficulty of accurately determining N₂O production [43].

In earlier studies respiration of feed was considered to have a negligible effect on utilisation of dry mass as determined gravimetrically [44] and therefore on CO₂ production. Later studies suggested that respiration by plant leaves can be an important source of error in the calculation of insect feed intake using gravimetric methods [45] and can cause major errors in energy budget studies of plant-feeding insects [46]. Our reported CO₂ production includes the respiration of the feed (Table 6). The extremely high contribution to total CO₂ production by the substrate of *P. marginata* (92.5%) was most likely due to large amounts of fungal biomass observed in the mixed feed and substrate when insects were absent in the experiments aimed to obtain correction values for CO₂-production by the substrate. No fungal growth was apparent during the experiments on feeding *P. marginata* larvae, suggesting that the contribution of the substrate to total respiration during the experiment was much lower. We conclude that the interaction between actively feeding *P. marginata* larvae and the substrate suppressed fungal growth through either consumption by the beetle larvae [47] of fungal biomass or through unknown chemical or combined chemical/mechanical mechanisms. Such interactions hinder the application of realistic corrections for the contribution of feed and substrate to the total CO₂ production and thus to quantify the CO₂ production arising from insect metabolism separately.

For all other species the relative contribution of the feed to total CO₂ production was minor, varying between 1.3% and 3.6%. Although feed respiration did have an impact on production of CO₂, still the production of CO₂ is much higher for *L. migratoria* than for the other insect species. A likely explanation for this higher production of CO₂ is the 7°C higher temperature *L. migratoria* was kept at, as a difference of 10°C is expected to double CO₂ production. Furthermore, the comparatively high ADG of *L. migratoria* is expected to result in higher production of CO₂.

In one of the repetitions for *A. domesticus*, a lower ADG and increased mortality were observed. Excluding this repetition, the emission of CO₂ per kg BM decreased slightly (68 vs 71 g/kg), but the emission of CO₂ per kg mass gain changed considerably (918 vs 1468 g/kg). This difference can for a large part be explained by a decrease in ADG (from 9.0 to 7.2%). *Acheta domesticus* did not produce CH₄, but N₂O production doubled (from 0.1 to 0.2 mg/kg BM; 1.9 vs 5.3 mg/kg mass gain). The production of CO₂ eq. also increased (0.04 vs 0.05 g CO₂ eq. /kg BM and 0.57 vs 1.57 g/kg mass gain). It is well possible that the higher N₂O production measured was caused by saprophytic bacteria utilising the dead *A. domesticus* and producing N₂O [6]. Although we included this repetition in the results, it is not clear whether this represents the practical situation best.

Table 6. Calculated CO₂ production of provided feed for five insect species recalculated per kg of animal body mass.

Species	CO ₂ production (g)/kg BM of insect	Relative contribution
<i>Pachnoda marginata</i>	46.2	92.46%
<i>Tenebrio molitor</i>	2.2	3.58%
<i>Blaptica dubia</i>	0.4	2.31%
<i>Acheta domesticus</i>	0.9	1.34%
<i>Locusta migratoria</i>	3.3	3.04%

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Large differences in NH₃ emission have been reported for conventional livestock. Pigs for example emit 4.8–75 mg/kg BM/day [48,49,50], poultry 72–436 mg/kg BM/day [41,49,51] and cattle 14–170 mg/kg BM/day [49,52,53]. Several factors influence NH₃ emission, such as temperature, relative humidity, food type, moisture content, pH, wind speed, housing type, and substrate [54,55].

In the current experiment, a clear NH₃ emission pattern was found; higher amounts of NH₃ were emitted during daytime for *A. domesticus*, *L. migratoria* and *B. dubia*, than during nighttime. Day-night rhythms for NH₃ excretion have been documented for pigs [5] and are strongly correlated with activity levels [56]. Quantitatively the differences between day and night emission levels are small; 7–10% with a maximum difference of 25% [5]. In our study this relative difference was approximately 33%. In all cases NH₃ emission levels were higher during the daytime than during the night-time. For *L. migratoria* this is the active period, for the nocturnal *B. dubia* and *A. domesticus* it is not, indicating that a different, unknown variable might influence NH₃ emission patterns in these insects.

NH₃ concentrations in the outgoing air, and consequently calculated NH₃ emission, increased from day one to day three in *B. dubia* (1.57 to 4.29 mg/kg BM/day) and *A. domesticus* (2.46 to 8.01 mg/kg BM/day). This could indicate that NH₃ emissions might be underestimated due to the relatively short time frame of our experiments. For *L. migratoria* NH₃ emission did not increase between day 1 and day 3 (5.57 and 5.05 mg/kg BM/day), suggesting that NH₃ production was stable. This might be caused by the faeces of this species that, contrary to those of *B. dubia* or *A. domesticus*, dry quickly after defecation.

We conclude that *P. marginata* and *T. molitor* probably did not emit NH₃. Poultry deep litter systems [57] have higher NH₃ emission rates than battery systems [55], which is explained by the presence of substrate.

The presence of substrates for *P. marginata* and *T. molitor* in this study corresponded with lower NH₃ emissions. A possible explanation is that gas exchange in the container is inhibited by

the substrate and therefore less emission of NH₃ was measured. However, it could also be that these species produce less NH₃.

All insect species in this study produced much lower amounts of NH₃ (3.0 to 5.4 mg/kg BM/day for *A. domesticus*, *L. migratoria* and *B. dubia*) than conventional livestock (4.8–75 mg/kg BM/day for pigs and 14–170 mg/kg BM/day for cattle). Further research is needed to determine for which insect species and to what extent NH₃ emissions increase further when a longer time frame is used.

Conclusions

To the authors' knowledge, the study presented here is the first to report on both GHG and NH₃ emissions of edible insect species. An evaluation of the GHG emissions of edible insect species is most relevant when based on CO₂ eq. per kg of mass gain. In that way a comparison of the selected species with each other and with conventional livestock is based on a cost-benefit principle, in which the GHG production (environmental cost) is directly linked to food production (benefit). GHG emission of four of the five insect species studied was much lower than documented for pigs when expressed per kg of mass gain and only around 1% of the GHG emission for ruminants.

The measured NH₃ emission levels of all insect species in this experiment were lower than reported NH₃ emission levels for conventional livestock.

The ADG of all insect species in this study was higher than for conventional livestock, while CO₂ production expressed as g/kg mass gain was comparable or lower, which indicates higher feed conversion efficiencies for insects.

This study therefore indicates that insects could serve as a more environmentally friendly alternative for the production of animal protein from the perspective of GHG and NH₃ emissions. A complete lifecycle analysis for species of edible insects is lacking at this point in time [58] and should be the focus point of further studies to allow a conclusive evaluation of the sustainability of insects as a protein-rich food source. The data presented in this study are indispensable for conducting a lifecycle analysis for edible insects.

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Author Contributions

Conceived and designed the experiments: DO HvdB JvL AvH. Performed the experiments: DO JvL. Analyzed the data: DO JvL MJWH. Contributed reagents/materials/analysis tools: MJWH. Wrote the paper: DO HvdB JvL AvH.

References

- IPCC (2007) Summary for Policymakers. In: Climate Change 2007: The Physical Science Basis, Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change Cambridge, United Kingdom and New York, NY, USA: IPCC.
- Steinfeld H, Gerber P, Wassenaar T, Castel V, Rosales M, et al. (2006) Livestock's long shadow; environmental issues and options Rome: Food and Agriculture Organization of the United Nations. 414 p.
- Kroon IJM, Holtslag AAM, Krol MC (2009) Inleiding atmosfeer (in Dutch). Wageningen: Wageningen University.
- de Vries M, de Boer IJM (2010) Comparing environmental impacts for livestock products: A review of life cycle assessments. *Livestock Science* 128: 1–11.
- Aarnink AJA, Keen A, Metz JHM, Speelman L, Verstegen MWA (1995) Ammonia emission patterns during the growing periods of pigs housed on partially slatted floors. *Journal of Agricultural Engineering Research* 62: 105–116.
- Wrage N, Velthof GL, van Beusichem ML, Oenema O (2001) Role of nitrifier denitrification in the production of nitrous oxide. *Soil Biology & Biochemistry* 33: 1723–1732.
- Fabbri C, Valli L, Guarino M, Costa A, Mazzotta V (2007) Ammonia, methane, nitrous oxide and particulate matter emissions from two different buildings for laying hens. *Biosystems Engineering* 97: 441–455.
- Lacey RE, JSRedwine, Parnell CB (2002) Emission factors for broiler production operations: A stochastic modeling approach. 2002 ASAE Annual International Meeting/CIGR XVth World Congress Chicago, Illinois.
- Cole NA, Clark RN, Todd RW, Richardson CR, Gueye A, et al. (2005) Influence of dietary crude protein concentration and source on potential ammonia emissions from beef cattle manure. *J Anim Sci* 83: 722–731.

10. Thorne PS (2007) Environmental health impacts of concentrated animal feeding operations: Anticipating hazards - Searching for solutions. *Environmental Health Perspectives* 115: 296–297.
11. Beets WC (1997) The need for an increased use of small and mini-livestock in integrated smallholder farming systems. *Ecology of Food and Nutrition* 36: 237–245.
12. Gullan PJ, Cranston PS (2005) *The insects: an outline of entomology*: Blackwell Publishing, pp 10–20.
13. Nakagaki BJ, DeFoliart GR (1991) Comparison of diets for mass-rearing *Acheta domesticus* (Orthoptera: Gryllidae) as a novelty food, and comparison of food conversion efficiency with values reported for livestock. *Journal of Economic Entomology* 84: 891–896.
14. Ramos-Elorduy J (2008) Energy supplied by edible insects from Mexico and their nutritional and ecological importance. *Ecology of Food and Nutrition* 47: 280–297.
15. Bruins E (2001) Geillustreerde Terrariumencyclopedie; Bruins E, ed. Lisse, The Netherlands: Rebo Productions BV.
16. Friederich U, Volland W (2004) *Breeding Food Animals: Live Food for Vivarium Animals*; Friederich U, Volland W, eds. Malabar, Florida: Krieger publishing company.
17. Verstegen MWA, Van Der Hel W, Brandsma HA, Henken AM, Bransen AM (1987) The Wageningen respiration unit for animal production research: A description of the equipment and its possibilities In: Verstegen MWA, Henken AM, eds. *Energy metabolism in farm animals: Effects of housing, stress and disease*. Dordrecht: Martinus Nijhoff Publishers, pp 21–48.
18. Wheeler EF, Casey KD, Zajackowski JS, Topper PA, Gates RS, Xin H, Liang Y, Tanaka A. Ammonia emissions from U.S. poultry houses: Part III - Broiler houses. 2003; Research Triangle Park, NC.
19. Terblanche JS, Chown SL (2007) The effects of temperature, body mass and feeding on metabolic rate in the tsetse fly *Glossina morsitans centralis*. *Physiological Entomology* 32: 175–180.
20. Bailey CG, Singh NB (1977) Energy budget for *Mamestra configurata* (Lepidoptera-Noctuidae). *Canadian Entomologist* 109: 687–693.
21. Emekci M, Navarro S, Donahaye E, Rindner M, Azrieli A (2004) Respiration of *Rhyzopertha dominica* (F.) at reduced oxygen concentrations. *Journal of Stored Products Research* 40: 27–38.
22. Gouveia SM, Simpson SJ, Raubenheimer D, Zanotto FP (2000) Patterns of respiration in *Locusta migratoria* nymphs when feeding. *Physiological Entomology* 25: 88–93.
23. Aidley DJ (1976) Increase in Respiratory Rate during Feeding in Larvae of Armyworm, *Spodoptera exempta*. *Physiological Entomology* 1: 73–75.
24. Armstrong G, Mordue W (1985) Oxygen consumption of flying locusts. *Physiological entomology* 10: 353–358.
25. Greenlee KJ, Harrison JF (2004) Development of respiratory function in the American locust *Schistocerca americana* I. Across-instar effects. *J Exp Biol* 207: 497–508.
26. Emekci M, Navarro S, Donahaye E, Rindner M, Azrieli A (2002) Respiration of *Tribolium castaneum* (Herbst) at reduced oxygen concentrations. *Journal of Stored Products Research* 38: 413–425.
27. Duncan FD, Krasnov B, McMaster M (2002) Metabolic rate and respiratory gas-exchange patterns in tenebrionid beetles from the Negev Highlands, Israel. *Journal of Experimental Biology* 205: 791–798.
28. May ML, Pearson DL, Casey TM (1986) Oxygen-Consumption of Active and Inactive Adult Tiger Beetles. *Physiological Entomology* 11: 171–179.
29. Kleiber M (1961) *The fire of life: an introduction to animal energetics* New York, London: John Wiley & Sons, Inc.
30. Agutter PS, Wheatley DN (2004) Metabolic scaling: Consensus or controversy? *Theoretical Biology and Medical Modelling* 1.
31. Heusner AA (1982) Energy metabolism and body size I. Is the 0.75 mass exponent of Kleiber's equation a statistical artifact? *Respiration Physiology* 48: 1–12.
32. da Silva JKL, Garcia GJM, Barbosa LA (2006) Allometric scaling laws of metabolism. *Physics of Life Reviews* 3: 229–261.
33. Lighton JRB, Fielden LJ (1995) Mass scaling of standard metabolism in ticks - a valid case of low metabolic rates in sit-and-wait strategists. *Physiological Zoology* 68: 43–62.
34. Chown SL, Marais E, Terblanche JS, Klok CJ, Lighton JRB, et al. (2007) Scaling of insect metabolic rate is inconsistent with the nutrient supply network model. *Functional Ecology* 21: 282–290.
35. Prothero JW (1979) Maximal Oxygen-Consumption in Various Animals and Plants. *Comparative Biochemistry and Physiology a-Physiology* 64: 463–466.
36. Collavo A, Glew RH, Huang Y-S, Chuang L-T, Bosse R, et al. (2005) House cricket small-scale farming. In: Paoletti MG, ed. *Ecological Implications of Minilivestock*: Science publishers, INC. pp 519–544.
37. Slansky FJ (1985) Food utilization by insects: Interpretation of observed differences between dry weight and energy efficiencies. *Entomologia Experimentalis et Applicata* 39: 47.
38. Hackstein JH, Stumm CK (1994) Methane production in terrestrial arthropods. *Proceedings of the National Academy of Sciences of the United States of America* 91: 5441–5445.
39. Eger M, Wagner B, Lemke T, Brune A, Friedrich MW (2003) Microbial community structure in midgut and hindgut of the humus-feeding larva of *Pachnoda ephippiata* (Coleoptera: Scarabaeidae). *Applied and Environmental Microbiology* 69: 6659–6668.
40. Guiziou F, Béline F (2005) In situ measurement of ammonia and greenhouse gas emissions from broiler houses in France. *Bioresource Technology* 96: 203–207.
41. Wathes CM, Holden MR, Sneath RW, White RP, Phillips VR (1997) Concentrations and emission rates of aerial ammonia, nitrous oxide, methane, carbon dioxide, dust and endotoxin in UK broiler and layer houses. *British Poultry Science* 38: 14–28.
42. Chadwick DR, Sneath RW, Phillips VR, Pain BF (1999) A UK inventory of nitrous oxide emissions from farmed livestock. *Atmospheric Environment* 33: 3345–3354.
43. Lemke R, Goddard T, Hahn D, Burton D, Ellert B, et al. (2002) An inter-laboratory comparison of nitrous oxide analysis in western Canada. *Communications in Soil Science and Plant Analysis* 33: 2705–2713.
44. Waldbauer GP (1968) The consumption and utilization of food by insects. *Advances in insect physiology* 5: 229–288.
45. Axelsson B, Ågren GI (1979) A correction for food respiration in balancing energy budgets. *Entomologia Experimentalis et Applicata* 25: 260–266.
46. van Loon JJA, Casas J, Pincebourde S (2005) Nutritional ecology of insect plant-interactions: persistent handicaps and the need for innovative approaches. *Oikos* 108: 194–201.
47. Li XZ, Brune A (2007) Transformation and mineralization of soil organic nitrogen by the humivorous larva of *Pachnoda ephippiata* (Coleoptera: Scarabaeidae). *Plant and Soil* 301: 233–244.
48. Cabaraux J-F, Philippe F-X, Laitat M, Canart B, Vandenheede M, et al. (2009) Gaseous emissions from weaned pigs raised on different floor systems. *Agriculture, Ecosystems & Environment* 130: 86–92.
49. Groot Koerkamp PWG, Metz JHM, Uenk GH, Phillips VR, Holden MR, et al. (1998) Concentrations and emissions of ammonia in livestock buildings in Northern Europe. *Journal of Agricultural Engineering Research* 70: 79–95.
50. Nicks B, Laitat M, Vandenheede M, Desiron A, Verhaeghe C, et al. (2003) Emissions of ammonia, nitrous oxide, methane, carbon dioxide and water vapor in the raising of weaned pigs on straw-based and sawdust-based deep litters. *Animal Research* 52: 299–308.
51. Demmers TGM, Burgess LR, Short JL, Phillips VR, Clark JA, et al. (1999) Ammonia emissions from two mechanically ventilated UK livestock buildings. *Atmospheric Environment* 33: 217–227.
52. Demmers TGM, Phillips VR, Short LS, Burgess LR, Hoxey RP, et al. (2001) SE—Structure and Environment: Validation of ventilation rate measurement methods and the ammonia emission from naturally ventilated dairy and beef buildings in the United Kingdom. *Journal of Agricultural Engineering Research* 79: 107–116.
53. Harper LA, Flesch TK, Powell JM, Coblenz WK, Jokela WE, et al. (2009) Ammonia emissions from dairy production in Wisconsin. *J Dairy Sci* 92: 2326–2337.
54. Casey JW, Holden NM (2006) Greenhouse gas emissions from conventional, agri-environmental scheme, and organic Irish suckler-beef units. *Journal of Environmental Quality* 35: 231–239.
55. Faulkner WB, Shaw BW (2008) Review of ammonia emission factors for United States animal agriculture. *Atmospheric Environment* 42: 6567–6574.
56. Blanes-Vidal V, Hansen MN, Pedersen S, Rom HB (2008) Emissions of ammonia, methane and nitrous oxide from pig houses and slurry: Effects of rooting material, animal activity and ventilation flow. *Agriculture, Ecosystems & Environment* 124: 237–244.
57. Volkova VV, Bailey RH, Wills RW (2009) *Salmonella* in broiler litter and properties of soil at farm location. *PLoS ONE* 4: e6403.
58. De Boer IJM (2008) Personal Communication. .
59. Beauchemin KA, McGinn SM (2005) Methane emissions from feedlot cattle fed barley or corn diets. *Journal of Animal Science* 83: 653–661.